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MODELLING EMISSIONS FROM PRESCRIBED BURNING USING FULLCAM

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Australian Government Department of Industry, Innovation and Science

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MODELLING CARBON EMISSION
*MADELLING CARBON EMISSION***S | REPORT NO. 479.2019**

TABLE OF CONTENTS

We would like to thank Dr Keryn Paul (CSIRO Land and Water) for sharing her latest research on the FullCAM fire module. We also would like to thank the Australian Government, Department of the Environment and Energy, Canberra for providing access to the latest research edition of the FullCAM software.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 **ABSTRACT**

The Full Carbon Accounting Model (FullCAM) is a software tool developed by the Australian Government, Department of the Environment and Energy as a s tandard method for carbon acco unting. It is primarily used as a means to report national greenhouse gas dynamics from the land sector due to anthropogenic activities. This study assessed the accuracy and usefulness of FullCAM in determining the mass of carbon (C) emissions produced from prescribed burning.

FullCAM proved to be a simple and reas onably reliable method fo r estimating C emissions from prescribed burni ng activities and for tr acking recovery of C pools related to forest ecosystems. In addition, C emissions from different prescribed burning scenarios and from wildfire can be easily compared. The FullCAM model can be used by land managers as a means to manage an important aspect of risk associated with planned burning. If land managers are required to perform C accounting activities in the future, the adoption of FullCAM will enable them to be c ompatible with the national standard of carbon accounting.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 **END-USER STATEMENT**

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The capacity to accurately predict carbon (C) emissions produced during prescribed burning activities is essential for desi gning regional burning programs ai med at optimising trade-offs among outcomes invo lving risk reduction and multifaceted ecosystem services including C storage.

Investigating the pote ntial of mo dels such as FullCAM to predi ct C emi ssions from prescribed burning in a cost- a nd time-effective way can potenti ally simplify the planning process i nvolved in designing such programs. It can also supp ort and improve the decision making required of burn planners to incorporate the effects of prescribed burning on C w ithin each burn block o ver the short- to medium-term. Testing the predictive capacity and limitations of FullCAM and its practicality for fire managers to use during planni ng can also be used to identify and address gaps in the model.

The initial investigation described here suggests that the publicly-available version of FullCAM can be u sed to pr edict emissions and simulate post-fire fuel loads with moderate accuracy. It is important to highlight that, in most cases for bark and in all cases for litter, FullCAM underpredicted post-fire biomass of these components when compared to field data. The discrepancies found between predicted and analysed data are discussed in this report and will be investigated further.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019
*NUMBER LING CARBON EMISSIONS | REPORT NO. 479.2019***</mark> 1. INTRODUCTION**

Empirical evidence collated over time has shown that prescribed burning can reduce the incidence and intensity of unplanned fires in the Australian landscape (Boer *et al*. 2009). However, due to increasing interest in evaluating the environmental impacts of prescribed burning, the importance of assessing carbon (C) dynamics and estimation of emissions from pl anned and unplanned fire is increasing. There are several approaches that la nd managers can use to assess carbon emissions based on empirical models that use f ield-collected data and/or process-based si mulation modelling.

While the use of empirical data will provide land managers wit h more a ccurate estimates, this method requires extensive collection of field data. Conversely, processbased simulation models allow field-collated datasets to be used in different scenarios to estimate various C pools in above- and belowground biomass, before and after fire, and C losses (emissions) as a result of fire and subsequent biomass recovery. The Full Carbon Accounting Model (FullCAM) is a software tool developed by the Australian Government, Department of the Environment and Energy that has been used as a means to report National greenhouse gas dynamics from the land sector due to anthropogenic activities (Department of the Environment and Energy 2016). A detailed description of the development of FullCAM has been reported b y Richards (2001) and Richards and Evans (2004). This study assessed the usefulness of FullCAM in determining the mass of emissions produced from prescribed burning and use of this tool by fire and land management agen cies when planni ng prescribed burning activities.

FullCAM uses a mass balance approach for C accounti ng. It is comprised of submodels that track changes in vegetation growth and litter decomposition, two processes intrinsically related to fuel accumulation, and soil C dynamics. FullCAM can be used to estimate C fluxes during undisturbed forest growth and the transition of plant components to debris (litter or surface fuels) and soil, and to predict regrowth of trees after disturbance such as harvesting and fire.

Models that are used to estimate emissions from the land sector can be categorised into three broad tiers based on their simplicity, the use of country-specific datasets and parameters, and the use of spatially-explicit datasets (Penman et al. 2003). FullCAM is classified as a Ti er 3 model as it provides more accurate and reli able estimates than Tier 1 and 2 models. This model uses spatially-explicit, fine resolution soil datasets (e.g. soil C and its fractions) and spatial-temporal climatic datasets available at 1 km spatial resolution (Department of the Environment and Energy 2016). Inputs or model drivers used in FullCAM can be readily added or adjusted. Changes in C stocks (above- and belowground) and atmospheric emissions are estimate d on a continuous basis (generally at monthly intervals) using non-li near processes that consider interactions among climate, soil and plant growth characteristics and land management activities. This contrasts to linear approaches that are commonly used in Tier 1 and 2 models.

,,,,,,,,,,,,,,,,,,, **1.1 OVERVIEW OF THE FULLCAM MODELLING FRAMEWORK**

FullCAM is a collection of sub-models that have been integrated to track C flows and emissions from agricultural systems (cropping and pasture) and forests (natural and managed). As this study used the FullCAM model for forests, a greater deal of attention will be directed towards the 'forest' suite of sub-models. FullCAM also has the capacity to incorporate 'land transitions' such as deforestation and reforestation. Brief descriptions of the main sub-models that mak e up the FullCAM model are provided in Table 1 and a depiction of the sub-models for C flow in both forest and agricultural systems are represented in Fig. 1.

TABLE 1. DESCRIPTION OF THE MAIN SUB-MODELS WITH THE FULLCAM MODEL. C = CARBON.

FIGURE 1. FLOW OF CARBON FOR SUB-MODELS IN FULLCAM INCLUDING THE AGRICULTURAL SYSTEM SUB-MODEL (LEFT) AND THE FOREST SYSTEM SUB-MODEL (RIGHT) (SOURCE: RICHARDS AND EVANS 2000A, B).

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 ,,,,,,,,,,,,,,,,,,,,,,, **1.2 TREE GROWTH MODELLING USING FULLCAM**

FullCAM adopts a hybrid approach for predicting the accumulation of aboveground biomass in woody vegetation at a given location. This approach requi res development of relationships between a long-term average process-based output called 'Forest Productivity Index' (FPI; a dimensionless index) that is derived using the 3-PG forest growth model and fi eld measurements of the maximum aboveground biomass from a forest stand that has had minimal disturbance (Landsberg and Waring 1997; Kesteven et al. 2004; Richards and Brack 2004). This empirical relationship is used to predict the parameter M (maximum aboveground biomass) for a given location using the FPI raster layer described as: The aboveground biomass of woody vegetation is predicted using a tree y ield formula (Waterworth et al. 2007) and is calculated as:

$$
M = (6.011 \times \sqrt{FPI} - 5.291)^{2}
$$
 Equation 1

where k is a stand constant which reflects the age of the maximum current annual increment, G is the age of maximum growth (in years), AGB is aboveground tree mass (t DM ha-1), r and y are tree y ield multipliers, M is maximum aboveground biomass (Equation 1) and d is the forest age (years) that can be adjusted to reflect different management actions.

Once aboveground biomass is estimated it is partitioned into various defined components using speci es-specific allocation tables (i.e. stem, branches and bark (aboveground) and coarse and fine roots (belowground)).

$$
k = 2G - 1.25
$$
 Equation 2.1

$$
AGB(t) = rMy \cdot e^{-k/d}
$$
 Equation 2.2

where k is a stand constant which reflects the age of the maximum current annual increment, G is the age of maximum growth (in years), AGB is aboveground tree mass (t DM ha-1), r and y are tree yield multipliers, M is maximum aboveground biomass (Equation 1) and d is the forest age (years) that can be adjusted to reflect different management actions.

Once aboveground biomass is estimated it is partitioned into various defined components using speci es-specific allocation tables (i.e. stem, branches and bark (aboveground) and coarse and fine roots (belowground)).

1.3 FIRE MODELLING USING FULLCAM

Disturbances to forests, such as fire, are included in FullCAM as an 'event'. Generally, there are six types of fires in FullCAM: (i) forest fire, (ii) prescribed burning, (iii) site preparation – broadcast burn, (iv) site preparation – windrow and burn, (v) wildfire – trees killed, and (vi) wildfire – trees not killed. The event category referred to as 'forest fire' is used when it cannot be determined if trees are or will be 'killed' or not 'killed'. The event category referred to as 'prescribed burning' is used for a human-i nduced fire when no trees are killed and 'forest fire' or 'wildfire' is used for a natural event.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 ,,,,,,,,,,,,,,,,,,,,,

Using the publicly available version of FullCAM software (Version: 4.1.6 19417), the model has the ability to calculate: (a) emissions from the combustion of live and dead organic material, (b) transition of derived resistant C (char) to soils, and (c) regrowth of vegetation (Fig. 2).

FIGURE 1. PARAMETER SETTINGS FOR FULLCAM FOR PRESCRIBED BURNING EVENTS (SOURCE: FULLCAM SOFTWARE).

In FullCAM, vegetation regrowth only happens if the fire is not a clearing event. There are 36 parameters associated with a forest fire event (Fig. 3) and brief descriptions of these parameters are described in the Appendix, Table A1.

Recent efforts to develop fire-related modelling capacity in FullCAM has been led by CSIRO Land and Water and supported by the Australian Government, Department of Environment. For this study, we were given access to the most recent research edition of FullCAM to test our simulation modelling (under a license agreement) and, as such, the fire-related parameters that have been introduced were used (Fig. 3). One of the newest additions are 'standing dead' parameters which are important for wildfires and 'year to re-growth' for post-fire recovery of vegetation. For certain parts of the study, the most recently available public version of the model was used to ensure the information provided was relevant to the level of access available to End Users.

FIGURE 2. UPDATED PARAMETER SETTINGS FOR FU LLCAM FOR PR ESCRIBED BURNING EVENTS (SOURCE: FULLCAM SOFTWARE).

The parameters related to 'Affected Portion' are associated with the combustion ratios of the exiting masses (both living and non-living) to the atmosphere and to debris or inert soil (see Table A1 in the Appendix). The current default values for fire event-related parameters are deriv ed from previ ous research on native forest fires from southern Australia and south eastern Queensland (Gould and Cheney 2008). A detailed review of these fire-related parameters can be found in Surawki *et al*. (2012).

This study will test the ability of FullCAM to estimate C emissions from different fuel components as a result of prescribed burni ng using field collated data to populate the simulation files.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 **2. METHODS**

2.1. INITIALISATION OF FULLCAM USING FIELD-COLLECTED FUEL LOADS

One of the major challenges of this study was mapping the field-collated datasets in FullCAM. Data collected from sites in the Australian Capital Territory (ACT; three sites, 2015) and New South Wales (NSW; nine sites, 2015–2016) were used (see Gharun *et al*. 2015; 2017; 2018; Bell *et al*. 2018). A brief description of the sites and i gnition data included in the FullCAM plot file and field sampling dates are presented in Table 2.

2.2. BUILDING FULLCAM PLOT FILES

Using the location information for the study sites, FullCAM plot files were built using the publicly available version of FullC AM and were later imported into the research edition. All input datasets (e.g. soil, climate) and default parameters were downloaded through the FullCAM database. The 'Eucalyptus Low Open Forest' forest type was used to build FullCAM plot files.

TABLE 2. GENERAL DESCRIPTION OF THE STUDY SITES IN NEW SOUTH WALES (NSW) AND THE AUSTRALIAN CAPITAL TERRITORY (ACT) TREATED WITH PRESCRIBED BURNING AND USED FOR MODELLING WITH FULLCAM.

2.3. INITIAL TREE BIOMASS

Since FullCAM simulates the understorey vegetation as part of the total biomass rather than explicitly simulated as a separate pool, understorey and overstorey biomass data

were amalgamated to d etermine the total aboveground biomass value for constraining in FullCAM. 'Total Biomass' (*TB*) was therefore calculated as:

Total biomass =
$$
\left(\frac{\alpha}{\beta}\right) \times 100
$$
 Equation 3

where *α* is the total aboveground biomass (AGB) and *β* is the FullCAM allocation sum for aboveground biomass which includes fractions for stems, branches, bark and leaves.

Once the total biomass was calculated, masses for tree parts were allocated (Fig. 4). FullCAM required biomass inputs as dry matter (t DM ha-1). The total biomass calculated for tree components for sites sampled in NSW and the ACT is provided in Table A2 in the Appendix. In the absence of field measured data, biomass in different tree components can be derived by using a separate plot file where *M* (Equation 1) is a set value equal to the total *AGB*, then d etermining the ou tput of mass of the components at equilibrium, with the assumption that a mature forest is present. In such a case, the model will run for a long time (500 years) using long-term climate datasets and allow the model to come to an equilibrium state.

2.4. INITIAL TREE DEBRIS/LITTER

Under the 'Tree Debris' tab in FullCAM, information regarding coarse woody debris (CWD) and debris can be found. In this study, 'debris' represents litter or surface fuel and is quantified in t C ha⁻¹. It is important to note that CWD is referred to as deadwood in FullCAM.

The measured bi omass of C WD was converted to mass of C by assuming the C content was 50% (the default FullCAM C content value for CWD). The calculated mass of C in CWD is segregated in FullCAM into two fractions; decomposable and resistant. Based on the latest calibration work by CSIRO Land and Water, CWD was assigned in the model as 0% decomposable and 100% resistant plant material (Fig. 5) (Paul, 2018, pers. comm.). After these calculations, data were i nserted in to the 'Deadwoo d' allocation under debris parameter (Fig. 5). The values used are provided in Table A2 in the Appendix.

FIGURE 4. ALLOCATION OF TO TAL BIOMASS TO C OMPONENTS AS DRY MATTER (T DM HA-1) (SOURCE: FULLCAM SOFTWARE).

The litter biomass measured at the study sites only accounted for the aboveground litter. For 'Belowground litter' (an input required by FullCAM is the 'litter' generated by roots), initial values were kept as default. Measured 'leaf' and 'decomposable leaf' fractions were considered as litter mass. Litter biomass (t DM ha-1) measured in the field was converted to C mass with the assumption that the C content is 52% (the default FullCAM C content value for leaves). A second assumption made was that the field litter biomass data was composed of leaves. Field data for 'Other' and 'Twigs' (t DM ha-1) were amalgamated and considered as 'Bark' to be compatible with the fuel load set up for litter in FullCAM. The mass of C in bark was calculated using 49% (the default FullCAM C content value for bark). Finally, decomposable plant material and resistant plant material segregations were done assuming that 0% bark litter and 77% of leaf litter went to the 'decomposable' fraction whilst 100% a nd 23%, respectively, went to the 'resistant' fraction for 'bark' and 'leaf' litter allocations (Paul, 2018, pers. comm.; Fig. 5). The calculations are displayed in Table A3 in the Appendix.

FIGURE 5. UPDATED LITTER POOLS IN FULLCAM USING FIELD DATA (SOURCE: FULLCAM SOFTWARE).

2.5 NEAR SURFACE LIVE PLANT BIOMASS

As might expected, field data for the near surface plant material (t DM ha-1) was small in contrast to the FullCAM downloaded data, given the near surface live plants growing under a forest receives less light than near surface live plants in a grassland context. For the simulation, an assumption was made that the most common native grass type in the area were grown under these forests. Measured biomass for grasses were assumed at equilibrium and inserted to FullCAM under the growth tab, under the native grass species selection. Some values reported were too low to use (<0.01) as the system requires at least one decimal value. Therefore, measured values for near surface biomass were rounded to one decimal point and used in the FullCAM model (Table A2 in the Appendix).

2.6 DEVELOPING EVENTS IN FULLCAM

The prescribed fire events created and simulated in FullC AM were based on the reported ignition dates for the sites studied (Table 2). When a range of ignition dates were given for site, the mid-date was considered. Two types of fire-related events were introduced; a 'prescribed fire' event that controls the impact of fire on tree components and tree litter and a 'grass fire' event that controls the impact of fire on near surface fuels an d litter generated from grasses. In FullCAM, these C pools are tracked as separate entities in forest and agriculture systems.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019
*NUMBER LING CARBON EMISSIONS | REPORT NO. 479.2019***</mark> 2.7 SENSITIVITY ANALYSIS**

Key parameters associated with a prescribed burning event in FullCAM are CWD, bark and leaf litter. Each of thes e fuel co mponents are further subdi vided into decomposable and resistant C fractions. Because these components are the major parameters that govern emissions from prescribed burning activities it is important to determine the effect that small changes in each of the fractions might have on overall C emissions. To do this, a simple approach was developed to test the sensitivity of key model parameters associated with the amount of C emitted during prescribed fire.

Three representative sites were selected for the sensitivity analysis, two sites sampled in NSW, Haycock Trig (HT) and Joadja (JOD), and one sampled in the ACT, Long Pine (LP). A total of 90 simulations were run for each site considering \pm 10% of the default parameter values (Table 3). Amounts of total C emitted were re ported when one parameter was changed within this range keeping all other parameters constant. In this way, the impact of individual parameters on total C emitted could be assessed.

TABLE 3. KEY PARAMETERS RELATED TO PRESCRIBED BURNING IN FULLCAM. DPM = DECOMPOSABLE PLANT MATTER; RPM = RESISTANT PLANT MATTER.

2.8 TESTING DIFFERENT BURN SCENARIOS USING FULLCAM AND THE IMPACT ON CARBON EMISSIONS

Five different burn scenarios were tested for each of the sites listed in the Table 2. The scenarios involved reduction of surface fuel (bark and litter fuels) by 25, 50, 75 and 100%. In addition, the default wildfire parameter settings available in the FullCAM were simulated for each site (Fig. 6). Descriptive statistics are reported, and a linear mixed model was fitted using total C emitted as the r esponse variable and the different scenarios as fixed effect terms (treatments). In the linear mixed model, the three plots for each site was treated as random effect terms. Once the model was fitted, mean separation was done using a Tukey's pair-wise mean separation test to assess whether there was a significant impact on C emissions based on the different scenarios tested. Fitting of the linear mixed model was done using nlme package in R statistical programming language (Pinheiro *et al*. 2018).

FIGURE 6. DEFAULT SETTINGS FOR WILDFIRE IN FULLCAM (SOURCE: FULLCAM SOFTWARE).

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019
NUMBER LING CARBON EMISSIONS | REPORT NO. 479.2019 **3. RESULTS AND DISCUSSION**

3.1 REPRESENTATION OF MEASURED FUEL LOADS IN FULLCAM

As mentioned previously, FullCAM only includes one vegetation layer for woody trees. Therefore, the amalgamation of elevated and overstorey fuel loads represented the total aboveground biomass for compati bility with F ullCAM specifications. Since prescribed burning does not generally target live woody fuel loads, this assumption is considered reasonable (Jenkins *et al*. 2015). Measured dry matter for the near surface live fuel was included in the model and, although not an accurate reflection, was represented as a native perennial grass species. In case of CWD and litter, these are represented in FullCAM under 'forest debris'. Within this section, bark fuel load is amalgamated with 'twig' and 'other' fuel load as FullC AM currently does not have the capacity to consider twigs as a separate entity.

3.2 EFFECT OF PRESCRIBED BURNING ON FINE FUEL LOAD MODELLED USING FULLCAM

The simulation modelled using FullCAM showed that for sites in both NSW and the ACT, the highest emissions were associated with the litter layer (Fig. 7). This fuel layer was comprised of C WD, bark, twigs, leaf litter and cont ributions of litt er from t he grass species present. Individual plots (1, 2 or 3) from each sampling site emitted varying amounts of C. For example, plots in the site referred to as Cotter in the ACT showed the greatest variation in C loss from the litter layer after a prescribed burning (Fig. 7b). Despite emission from Plots 1 and 2 at this site being relatively similar (2.5–3.0 t C ha⁻¹), Plot 3 emitted approximately 8 t C ha⁻¹, almost four times more than emitted from Plots 1 and 2 (Fig. 7b). Other sites in the ACT (Googong, Lone Pine, Tidbinbilla) remained relatively consistent for each plot (Fig. 7b).

Emissions from the litter fuel layer for sites in NSW were relatively consistent among plots (Fig. 7a). The exception was Ma rtins Creek (MTC), where emissions from Plot 3 were lower (2 t C ha⁻¹) compared to Plots 1 and 2 (5–6 t C ha⁻¹). Influences such as initial biomass, fire intensity, fuel moisture content and position within the landscape would all have had an effect on how much fuel was actually burnt and, consequently, how much C was emitted.

The total C emitted was also simulated through FullCAM to visually depict the overall C loss across all plots at each site (Fig. 8). At two sites in NSW, Joadja (JOD) and Martins Creek (MTC), the greatest emissions were produced in one plot only at each site (Fig. 8a). For sites in the ACT, Plot 3 at Cotter produced the greatest amount of C lost to the atmosphere (Fig. 8b).

For sites in NSW, the mean emitted total C varied from 2.0–6.8 t C ha⁻¹ (Table 4). These values are similar to emission estimates calculated for C emitted after pres cribed burning in a previous study (Volkova and Weston 2013). Variation in emissions was low at most sites, except for at Joadja (JOD) and Martins creek (MTC) as previously noted. Mean emitted C values estimated for sites in the ACT did not show much variation in contrast to sites in NSW, however, the standard deviation for plots in Cotter (COT) was

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 ,,,,,,,,,,,,,,,,

much larger than for the o ther sites (Table 4). Once again, this shows the large variation of emissions at the plot level. Overall, mean values for emitted C for the four sites sampled in the ACT were higher compared to the nine sites sampled in NSW (4.20 M 0.80 SD and 3.67 M 1.64 SD \uparrow C ha⁻¹, respectively).

FIGURE 7. CARBON (C) EMISSION ESTIMATES (T C HA-1) FOR DIFFERENT FUEL COMPONENTS (LITTER, NEAR SURFACE, TREE) EMITTED DURING PRESCRIBED BURNING FOR STUDY SITES IN (A) NEW SOUTH WALES (NSW) AND (B) THE AUSTRALIAN CAPITAL TERRITORY. REFER TO TABLE 2 FOR NAMES OF SITES FROM NSW.

FIGURE 8. TOTAL CARBON (C) EMITTED (T C HA⁻¹) DUE TO PRESCRIBED BURNING FOR STUDY SITES IN (A) NEW SOUTH WALES (NSW) AND (B) THE AUSTRALIAN CAPITAL TERRITORY. REFER TO TABLE 2 FOR NAMES OF SITES FROM NSW.

TABLE 4. ESTIMATED AMOUNT OF CARBON (C) EMITTED (MEAN ± STANDARD DEVIATION) FOR SITES IN NEW SOUTH WALES (NSW) AND THE AUSTRALIAN CAPITAL TERRITORY (ACT).

,,,,,,,,,,,,,,,,,

3.3 ESTIMATION OF CARBON EMITTED ACROSS THE LANDSCAPE

Using the calculated mean C emission values and t he corresponding standard deviation values for respective sites, a simple approach was ado pted to calculate landscape-scale C emitted due to prescr ibed burning for the sites in NSW (Table 5). The upper and low er confidence intervals around the mean emitted C (95% confidence interval) were estimated using the burnt areas for the respective sites in NSW. These estimates provided a range over which we can be 95% certain contains the true mean. The calculated confidence intervals for all sites reported high variation due to limited number of sam ples for a giv en site. Nevertheless, reporting such uncertainties are essenti al for effective decision making, planning an d policy directions related to prescribed burning activities in fire and land management agencies.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 ,,,,,,,,,,,,,,,,,,, **TABLE 5.** UPPER AND LOWER CONFIDENCE INTERVALS OF ESTIMATED AMOUNT OF TOTAL CARBON (C)

EMITTED FROM STUDY SITES BURNT WITH PRESCRIBED FIRE IN NEW SOUTH WALES.

3.4 RECOVERY OF BIOMASS LOSS FROM TREES DUE TO PRESCRIBED BURNING

According to FullCAM, biomass lost from the woody trees was recovered within 2–3 years after the prescribed burni ng. Since prescribed burns do not target live woody compartments, the r ecovery rates simulated by FullCAM model are acceptable. Recovery of the live woody compartments in terms of aboveground biomass (t DM ha⁻¹) for the selected sites are shown in Fig. 9.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019
NATALA AND RELATION SANDS | REPORT NO. 479.2019

FIGURE 9. RECOVERY OF WOODY VEGETATION AFTER FUEL REDUCTION BURNING AT SITES IN NEW SOUTH WALES (NSW) AND THE AUSTRALIAN CAPILTAL TERRITORY (ACT) AS MODELLED BY FULLCAM. REFER TO TABLE 2 FOR NAMES OF SITES FROM NSW.

3.5 SENSITIVITY ANALYSIS

Changing the default parameter v alues associated with prescr ibed burning, had negligible effect generally on the average amounts of C emitted from each site and fuel component considered (Table 6). The standard deviation reported for HT was the same for all fuel components. For the other two sites, JOD and LP, values varied only slightly for each of the fuel components tested.

Changing the default parameters by 10% had little effect on total C emitted (Fig. 10). As expected from the small changes i n individual fuel components (Table 6), the distribution (mean ± standard deviation) of values was the same for HT and very similar for JOD an d LP. Litter, CWD and bark bi omass for HT was two to five times smaller compared to the other two sites (Table 7), suggesting that the sensitivity of the three parameters considered was mainly due to variation in fuel loads.

It should be noted that this sensitivity analysis was done by considering one parameter at a time. Use of advanced sampling schemes, such as conditional latin hypercube, using a much hi gher number of iterations ($n = 1000$) and running all par ameters simultaneously are recommen ded for a more det ailed sensitivity analysis. For example, Paul *et al*. (2013) assessed the sensitivity of the FullCAM model parameters using a similar framework.

TABLE 6. AMOUNTS OF CARBON (C) EMITTED AFTER VARYING (± 10%) KEY DEFAULT PARAMETERS IN FULLCAM RELATED TO PRESCRIBED BURNING. CWD = COARSE WOODY DEBRIS.

,,,,,,,,,,,,,,, **TABLE 7.** DERIVED DATA FOR COARSE WOODY DEBRIS (CWD), LITTER AND BARK FUEL FOR THE TWO TEST SITES SAMPLED IN NEW SOUTH WALES (HAYCOCK TRIG AND JOADJA) AND ONE TEST SITE SAMPLED IN THE AUSTRALIAN CAPITAL TERRITORY (LONE PINE). DM = DRY MATTER.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019

The emissions reported for the 90 simulations used in this analysis are available in Table A4 in the Appendix.

FIGURE 10. DISTRIBUTION (MEAN ± STANDARD DEVIATION) OF TOTAL CARBON EMISSIONS FOR THREE SITES WITH RESPECT TO CHANGING DEFAULT PARAMETERS BY ± 10% FOR BARK, COARSE WOODY DEBRIS (CWD) AND LITTER.

,,,,,,,,,, **3.6 CARBON EMISSIONS FROM DIFFERENT BURNING SCENARIOS AND WILDFIRE**

As expected, when the 'burn percentage' parameters of surface fuel (bark and litter) were increased from 25 to 100%, FullCAM simulated C emissions also increased (Fig. 11). For example, the mea n total C emissions from Cotter in the ACT was approximately 3 t C ha-1 for the scenario with 25% of the vegetation burnt and 5.4 t C ha⁻¹ for the scenario with 100% of the vegetation burnt (Fig. 11). This pattern was common for the all the sites considered in the s tudy. A wildfire scenario was also simulated for all sites (Fig. 12). Carbon emissions from simulated wildfire were up to 40 times greater compared to the prescribed burning scenarios. Variability in the standard error values associated with total C emitted from simulated wildfire at each site reflects site-specific variation in fuel loads (Fig. 12).

FIGURE 11. CARBON (C) EMISSIONS (MEAN ± STANDARD ERROR OF MEAN; T C HA-1) FOR FOUR BURN SCENARIOS VARYING SURFACE FUEL (BARK AND LITTER) AND KEEPING ALL OTHER MODEL PARAMETERS CONSTANT: (A) SCENARIO_25 = 25% OF BARK AND LEAF LITTER REMOVED; (B) SCENARIO_50 = 50% OF BARK AND LEAF LITTER REMOVED; (C) SCENARIO_75 = 75% OF B ARK AND LEAF LITTER REMOVED; (D) SCENARIO_100 = 100% OF BARK AND LEAF LITTER REMOVED. REFER TO TABLE 2 FOR NAMES OF SITES FROM NEW SOUTH WALES.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 ,,,,,,,,,,,,,,,,,

FIGURE 12. CARBON (C) EMISSIONS (MEAN ± STANDARD ERROR OF MEAN; T C H A⁻¹) FOR W ILDFIRE SIMULATED USING THE DEFAULT VALUES (FIG. 6) IN FULLCAM. REFER TO TABLE 2 FOR NAMES OF SITES FROM NEW SOUTH WALES.

The simulation analysis showed that for each of the prescribed burning s cenarios tested, the amounts of C emitted t o the atmosphere was hi ghly variable and site specific (Table 8). To assess whether the different burn scenarios had a s tatistically significant effect on the total amount of C emitted, linear mixed modelling and Tukey's paired-wise mean separation tests were used. A summary of the estimated fixed effect terms derived from the linear mixed models are presented in Appendix A5 (Summary of the linear mixed model an alysis). Each paired-wise comparison was statistically significant at 0.05 probability level (Table 9).

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 7. **TABLE 8.** TOTAL CARBON (C) EMISSIONS ESTIMATED FOR EACH SITE FOR THE SCENARIOS TESTED. SCENARIO_25 = 25% OF BARK AND LEAF LITTER REMOVED; SCENARIO_50 = 50% OF BARK AND LEAF LITTER REMOVED; SCENARIO_75 = 75% OF BARK AND LEAF LITTER REMOVED; SCENARIO_100 = 100% OF BARK AND LEAF LITTER REMOVED; SCENARIO_WF = SIMULATED WILDFIRE USING DEFAULT VALUES. REFER TO TABLE 2 FOR NAMES OF SITES FROM NEW SOUTH WALES.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019 ,,,,,,,,,,,,,,,,,,

TABLE 9. TUKEY'S PAIRED-WISE MEAN SEPARATION FOR TOTAL CARBON EMITTED UNDER DIFFERENT PRESCRIBED BURNING SCENARIOS. * = SIGNIFICANT AT 0.95% CONFIDENCE INTERVAL (0.05 PROBABILITY LEVEL); SCENARIO_25 = 25% OF BARK AND LEAF LITTER REMOVED; SCENARIO_50 = 50% OF BARK AND LEAF LITTER REMOVED; SCENARIO 75 = 75% OF B ARK AND LEAF LITTER REMOVED; SCENARIO 100 = 100% OF BARK AND LEAF LITTER REMOVED.

3.7 COMPARISON OF SIMULATED POST-FIRE BIOMASS FOLLOWING PRESCRIBED BURNING

To assess the quality of prediction estimates simulated with the FullCAM model, postfire fuel loads ('simulated'; total biomass, CWD, bark and litter) were compared with field observations ('measured') (Figs. 13–16). Overall, FullCAM was able to s imulate post-fire fuel loads wit h moderate accuracy. Me an values for si mulated and measured total aboveground biomass were similar for Googong, Haycock Trig (HT), Left Arm (LEF) and Lone Pine (Fig. 12) but were noticeably different for the other sites with examples of both under- and overprediction from FullCAM. For CWD, close matches between mean measured and simulated biomass values were found for HT, LEF, Joadja (JOD) and Martins Creek (MTC) (Fig. 14) and, again, for the remaining sites with evidence of both under- and overpredictions. For bark and litter, there were very poor relationships between model predictions and field data for all sites (Figs. 14 and 15). In most cases for bark and in all cases for litter, FullCAM underpredicted biomass of these two components when compared to field data.

The discrepancy between measured and simulated fuel loads may, in part, be due to the decision made in the initial stage of modelling to aggregate both o ver- and understorey biomass instead of treating both fractions as separate biomass pools. The requirement to use partitioning, decomposition and growth parameters relating to grasses as a substitute for woody sclerophyllous shrubs as understorey fuel is far more likely to be the reason for poor compari sons between measured and simulated fuel loads. This is a shortcoming of the FullCA M model that could be addressed in future reiterations of the model.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019
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FIGURE 13. COMPARISONS BETWEEN SIMULATED AND MEASURED TOTAL ABOVEGROUND BIOMASS (T DM HA⁻¹) AFTER PRESCRIBED BURNING AT EACH OF THE STUDY SITES. STUDY SITES IN NEW SOUTH WALES INCLUDE: HES = HELICOPTER SPUR; HT = HAYCOCK TRIG; JOD = JOADJA; KIF = KIEF TRIG; LAK = LAKESLAND; LEF = LEFT ARM; MTC = MARTINS CREEK; PTS = PATERSON; SG = SPRING GULLY; AND IN THE AUSTRALIAN CAPITAL TERRITORY INCLUDE: COTTER; GOOGONG; LONE PINE; TIDBINBILLA.

MODELLING CARBON EMISSIONS | REPORT NO. 479.2019
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FIGURE 14. COMPARISONS BETWEEN SIMULATED AND MEASURED COARSE WOODY DEBRIS BIOMASS (CWD; T DM HA-1) AFTER PRESCRIBED BURNING AT EACH OF THE STUDY SITES. STUDY SITES IN NEW SOUTH WALES INCLUDE: HES = HELICOPTER SPUR; HT = HAYCOCK TRIG; JOD = JOADJA; KIF = KIEF TRIG; LAK = LAKESLAND; LEF = LEFT ARM; MTC = MARTINS CREEK; PTS = P ATERSON; SG = SPRING GULLY; AND IN THE AUSTRALIAN CAPITAL TERRITORY INCLUDE: COTTER; GOOGONG; LONE PINE; TIDBINBILLA.

FIGURE 15. COMPARISONS BETWEEN SIMULATED AND MEASURED BARK BIOMASS (T DM HA⁻¹) AFTER PRESCRIBED BURNING AT EACH OF THE STUDY SITES. STUDY SITES IN NEW SOUTH WALES INCLUDE: HES = HELICOPTER SPUR; HT = HAYCOCK TRIG; JOD = JOADJA; KIF = KIEF TRIG; LAK = LAKESLAND; LEF = LEFT ARM; MTC = MARTINS CREEK; PTS = PATERSON; SG = SPRING GULLY; AND IN THE AUSTRALIAN CAPITAL TERRITOY INCLUDE: COTTER; GOOGONG; LONE PINE; TIDBINBILLA.

FIGURE 16. COMPARISONS BETWEEN SIMULATED AND MEASURED LITTER BIOMASS (T DM HA⁻¹) AFTER PRESCRIBED BURNING AT EACH OF THE STUDY SITES. STUDY SITES IN NEW SOUTH WALES INCLUDE: HES = HELICOPTER SPUR; HT = HAYCOCK TRIG; JOD = JOADJA; KIF = KIEF TRIG; LAK = LAKESLAND; LEF = LEFT ARM; MTC = MARTINS CREEK; PTS = PATERSON; SG = SPRING GULLY; AND IN THE AUSTRALIAN CAPITAL TERRITORY INCLUDE: COTTER; GOOGONG; LONE PINE; TIDBINBILLA.

4. CONCLUSIONS

We have demonstrated a cost - and time-effective, simple approach to deriving estimates of C emissions from prescribed burning. This simple approach will assist land managers to estimate C emissions associated with prescribed burning that can be used for multiple planning activiti es, such a s to manage bushfire risk and carbon dynamics. While we acknowledge that FullCAM has limitations in modelling C emission estimations, it has the capacity to track carbon pools related to forest ecosystems. As FullCAM is used in Australia for the national standard of carbon accounting, it is subject to continuous ongoing development, with improved model versions released through the Department of Environment and Energy. The use of FullCAM by fire and land management agencies will enable them to be compatible with the national standard of carbon acco unting and agen cies will be prepared for any carbon accounting activities that are required in their practice.

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6. APPENDIX

TABLE A1. GENERAL DESCRIPTION OF FIRE EVENT-RELATED PARAMETERS WITHIN FULLCAM (FULLCAM HELP).

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TABLE A2. DERIVED DATA FOR TREE COMPONENTS AND COARSE WOODY DEBRIS FOR SITES IN NEW SOUTH WALES (NSW) AND THE AUSTALIAN CAPITAL TERRITORY (ACT) AS REQUIRED BY FULLCAM. THE UNIT FOR ALL BIOMASS SAMPLES IS T DM HA-1. CWD = COARSE WOODY DEBRIS; DM = DRY MATTER; LAT. = LATITUDE; LONG. = LONGITUDE; AGB = ABOVEGROUND BIOMASS; BGB = BELOWGROUND BIOMASS; C = CARBON; RPM = RESISTANT PLANT MATTER. STUDY SITES IN NSW INCLUDE; HT = HAYCOCK TRIG; SG = SPRING GULLY; HES = HELICOPTER SPUR; PTS = PATERSON; LAK = LAKESLAND; LEF = LEFT ARM; JOD = JOADJA; MTC = MARTINS CREEK; KIF = KIEF TRIG; AND IN THE ACT INCLUDE: GOO = GOOGONG; TID = TIDBINBILLA; LP = LONE PINE; COT = COTTER.

Data measured in field/site location Calculated values for the FullCAM model

TABLE A3. DERIVED DATA FOR LITTER FOR SITES IN NEW SOUTH WALES (NSW) AND THE AUSTRALIAN CAPITAL TERRITORY (ACT) AS REQUIRED BY THE FULLCAM MODEL. DM = DRY MATTER; C = CARBON; DPM = DECOMPOSABLE PLANT MATTER; RPM = RESISTANT PLANT MATTER. STUDY SITES IN NSW INCLUDE: HT = HAYCOCK TRIG; SG = SPRING GULLY; HES = HELICOPTER SPUR; PTS = PATERSON; LAK = LAKESLAND; LEF = LEFT ARM; JOD = JOADJA; MTC = MARTINS CREEK; KIF = KIEF TRIG; AND IN THE ACT INCLUDE: GOO = GOOGONG; TID = TIDBINBILLA; LP = LONE PINE; COT = COTTER.

Mean values were calculated using pseudo replicates for a given site/plot using measured datasets Calculated values to insert into the FullCAM model

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TABLE A4. SUMMARY OF CARBON EMITTED FOR 90 SIMULATIONS USING DATA FROM PRESCRIBED FIRES IN THREE SITES IN NEW SOUTH WALES AND THE AUSTRALIAN CAPITAL TERRITORY. CWD = COARSE WOODY DEBRIS.

A5. SUMMARY OF THE LINEAR MIXED MODEL ANALYSIS

Linear mixed-effects model fit by REML Data: stacked_all AIC BIC logLik 371.0246 389.1679 -179.5123 Random effects: Formula: ~1 | Source (Intercept) Residual
StdDev: 1.441097 0.4776447 1.441097 0.4776447 Fixed effects: Total_C_mass_emitted ~ Fire Value Std. Error DF t-value p-value (Intercept) 4.309696 0.2431052 114 17.727701 0 FireScenario_25 -2.343778 0.1081652 114 -21.668499 0 Fi reScenari o_50 -1.562519 0.1081652 114 -14.445666 0 FireScenario_75 -0.781259 0.1081652 114 -7.222833 0 Correlation: (Intr) FrS_25 FrS_50 Fi reScenari o_25 -0.222 FireScenario_50 -0.222 0.500 FireScenario_75 -0.222 0.500 0.500 Standardized Within-Group Residuals: Min Q1 Med Q3 Max -3.340618482 -0.402741244 -0.001257532 0.369200929 3.704903609 Number of Observations: 156 Number of Groups: 39

